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MILFP Model and Algorithms for Network Design and Long-Term Planning of Water Management System for Shale Gas Production

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This paper addresses the optimal design and operations of water supply chain networks for shale gas production. We develop a mixed-integer linear fractional programming (MILFP) model with the objective to maximize profit per unit freshwater consumption, such that both economic performance and water-use efficiency are optimized. The model simultaneously accounts for the design and operational decisions for freshwater source selection, multiple transportation modes, and water management options. Water management options include underground injection, commercial centralized wastewater treatment (CWT), and different onsite treatment technologies. To globally optimize the resulting MILFP problem efficiently, we present three tailored solution algorithms: a parametric approach, a reformulation-linearization method, and a novel Branch-and-Bound & Charnes-Cooper transformation method. The proposed models and algorithms are illustrated through one case study based on Marcellus shale play, in which onsite treatment shows its superiority in improving freshwater conservancy, maintaining a stable water flow, and reducing transportation burden.

1. Introduction

Natural gas is playing a significant role in meeting global energy demand, and is also serving as a transition fuel as the U.S. develops more sustainable fuel options. Shale gas is unconventional natural gas extracted from shale rock and has emerged as one of the most promising energy sources within the past decade. In 2012, 35% of the U.S. natural gas production was from shale gas (EIA, 2011). With increasing production of shale gas, the U.S. has changed from an importer to a net exporter of natural gas (EIA, 2013). The recent large-scale production of shale gas would not have been possible without the development of hydraulic fracturing and horizontal drilling technologies (Gregory et al., 2011).

Despite the economic potential of using hydraulic fracturing and horizontal drilling technologies for shale gas production, there are increasing concerns about its environmental impacts (Kotek and Tabasb, 2013). In 2006, about 35,000 shale wells were drilled in the U.S., and each well required approximately 4-6 million gallons of injected water for shale gas production (Jiang et al., 2013). Meanwhile, in shale plays such as Marcellus, 10%-25% of the injected water flows back to the surface as highly contaminated water. This water contains a high concentration of total dissolved solids (TDS) as well as other toxic and radioactive dissolved constituents (Karapataki, 2012), which is challenging and costly to treat. Economical production of shale gas requires effective wastewater management to minimize freshwater consumption while ensuring sufficient water supply to fracturing operations. Due to different water flow rates and water compositions in the shale wells, it is very important to determine the corresponding optimal strategies for water management.

Water from freshwater sources is transported to shale sites by pipelines or trucks. At shale sites, fracturing fluid is prepared and pumped into the wellbore at a high pressure, meanwhile a certain amount of the injected fracturing fluid returns to the surface as flowback water. This flowback water can be classified by the

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concentration of TDS (Slutz et al., 2012). The resulting wastewater can be temporarily stored in tanks or impoundments, transported to Class II disposal wells for underground injection, transported to commercial centralized wastewater treatment (CWT) facilities for treatment, or directly treated by onsite treatment facilities for reuse (Veil, 2010). Multiple technologies are involved in each of these water management options.

In this work, we propose a novel mixed-integer linear fractional programming (MILFP) model for the optimal design and operations of water supply chain networks for shale gas production. We consider a fractional function as the objective, the numerator of which is the profit for shale gas production and the denominator is the net freshwater consumption. This objective function reflects the profit associated with unit net consumption of freshwater. The model takes into account both strategic design and operational planning decisions of water supply chain networks for shale gas production. To facilitate the solution process of resulting MILFP problem, we present three tailored global optimization algorithms: a parametric algorithm based on the exact Newton's method, a reformulation-linearization method, and an algorithm integrating the Branch-and-Bound and Charnes-Cooper transformation methods. One case study based on Marcellus shale play is given to demonstrate the proposed optimization model and algorithms.

2. Model Formulation

In this section, we present the MILFP model to address the optimal design and operations of water supply chain networks for shale gas production. A mixed-integer linear programming (MILP) model is also presented. The MILFP model (P) maximizes the profit per unit freshwater consumption, which is subject to six types of constraints: mass balance constraints, cost constraints, capacity constraints, composition constraints, lower and upper bounding constraints, and logic constraints. The MILP model (EP) minimizes the total cost of this water supply chain network, which is subject to the same constraints as in MILFP model (P).

MILFP problem (P)		MILP problem (EP)	
$\max npf = \frac{NP - cw}{nf}$	(1)	min <i>cw</i> given in equation (8	8)
s.t. mass balance constraints	(2)	s.t. mass balance constraints (9)
cost constraints	(3)	cost constraints (10)
capacity constraints	(4)	capacity constraints (11)
composition constraints	(5)	composition constraints ((12)
bounding constrains	(6)	bounding constrains	(13)
logic constrains	(7)	logic constrains	(14)

The term NP stands for the total net present revenue gained by shale gas production, which is calculated by,

$$NP = \sum_{i \in I} \sum_{j \in J} \sum_{l \in L} \sum_{i \in T} \frac{SP_{i,t} \cdot WP_{i,j,l,t} \cdot CC_{i,j,t}}{(1 + DR)^{t}}$$
(15)

where $SP_{i,t}$ denotes the average revenue per unit shale gas production at shale site i at time period t, $WP_{i,j,l,t}$ denotes the TDS concentration range *I* wastewater production profile for well *j* at shale site *i* at time period *t*, $CC_{i,i,i}$ is the correlation coefficient between water and shale gas production profiles for well *i* at shale site *i* at time period *t*, and *DR* is the discount rate per time period.

The term *cw* stands for the total net present cost in the water supply chain, including the following items:

$$cw = c_{water} + c_{transport} + c_{handling}$$
(16)

where cwater denotes the total net present cost for freshwater acquisition; ctransport denotes the total net present cost for water transportation, including both freshwater and wastewater; and ctreatment denotes the total net present cost for handling wastewater by different options.

The term *nf* denotes the net freshwater consumption. Since the water treated by CWT facilities can be directly sent to surface discharge, the discharged water returns to the natural water cycle and does not count as net freshwater consumption. Thus, the net freshwater consumption is given by,

$$nf = \sum_{s \in S} \sum_{i \in I} \sum_{m \in M} \sum_{t \in T} fw_{s,i,m,t} - \sum_{c \in C} \sum_{t \in T} wtcd_{c,t}$$
(17)

3. Solution Approaches

In this section, we present three tailored global optimization algorithms for efficient solution of this MILFP problem, namely a parametric algorithm, a reformulation-linearization method, and a Branch-and-Bound algorithm that integrates with Charnes-Cooper transformation (Charnes and Cooper, 1962).

3.1 Parametric algorithm

The main idea of this algorithm is to transform the original MILFP problem into an equivalent parametric MILP problem F(q). This problem has the same constraints but a different objective function formulated as the numerator of the original objective function minus the denominator multiplied by a parameter q. When F(q)=0, the inner MILP problem has a unique optimal solution which is exactly the same as the global optimal solution to the original MILFP problem (You et al., 2009). Therefore, solving the MILFP problem becomes equivalent to finding the root of the equation F(q)=0. Though F(q) does not have a closed-form analytical expression, we can apply numerical root-finding approaches such as Newton's method to solve this problem (Zhong and You, 2014). In this work, we apply the parametric algorithm based on the exact Newton's method.

3.2 Reformulation-linearization algorithm

The reformulation-linearization approach integrates the Charnes-Cooper transformation with Glover's linearization scheme (Glover, 1975). The key idea is to transform the original MILFP problem into an exactly equivalent MILP problem by introducing auxiliary variables and constraints (Yue et al., 2013). To do this, we first apply Charnes-Cooper transformation to reformulate the original MILFP problem into an equivalent MINLP problem. Glover's linearization is then employed to convert bilinear terms and corresponding constraints into linear ones. Thus, an exactly equivalent MILP problem is obtained based on the MINLP problem, which can be solved efficiently using the Branch-and-Cut algorithm implemented in solvers like CPLEX. This approach only needs to solve the reformulated MILP once, but the reformulated MILP problem has a larger size than the original MILFP problem, which might need substantially more computational time (Yue et al., 2013).

3.3 Branch-and-Bound & Charnes-Cooper transformation algorithm

The Branch-and-Bound & Charnes-Cooper transformation algorithm integrates the Branch-and-Bound algorithm (Gupta and Ravindran, 1985) and the Charnes-Cooper transformation method (Charnes and Cooper, 1962). Since MILFP problems are a special class of MINLP problems, we are able to apply a similar Branchand-Bound algorithm for solving general MINLP problems to the global optimality of MILFP problems. The procedure of the proposed Branch-and-Bound & Charnes-Cooper transformation algorithm is introduced stepby-step in Table 1.

1:	Initialization.	15:	if the objective value > bestfound then
	Set bestfound = -Inf, bestpossible = +Inf		
2:	Relax the MILFP problem to obtain the relaxed	16:	Set bestfound = obi
	LFP problem		Set bestiound – obj
3:	Add node 1 to the waiting list	17:	Remove all the nodes in the waiting list
	Set upper bound of node 1 = +Inf		with their bound < bestfound
4:	for node n=1,2, in the waiting list, do	18:	end if
5:	if node <i>n</i> has the highest bound, then	19:	else
6:	Remove this node from the waiting list	20:	Select the variable w_j farthest from 0 and u
7:	Transform the LFP to an equivalent LP by	21:	Create two new nodes based on the current
	applying Charnes-Cooper transformation		one with $w < 0$ and $w > u$ respectively
	and introducing auxiliary variables		······································
8:	Import bounding constraints	22:	
	$w_i = 0 \{ j \in J y_i = 0 \}; w_i = u \{ j \in J y_i = 1 \}$		Store the new nodes in the waiting list
9:	Solve the LP subproblem	23:	end if
10:	if solver encounters an error then	24:	else if this subproblem is infeasible then
11:	Abort this iteration	25:	Abort and fathom this node
12.	also if a fassible solution is obtained then	26.	and if
12.		20.	
13:	If all the variables $w_j = 0$ or u , then	27:	end if
14:	Store as a new integer solution	28:	end for

Table 1. Pseudocode of the Branch-and-Bound & Charnes-Cooper transformation algorithm

4. Case studies

To illustrate the application of the proposed models and solution approaches, we present one case study based on Marcellus shale play. All the models and solution procedures are coded in GAMS 24.2.2. The MILP

problems are solved using CPLEX 12.6. The MINLP solvers utilized are DICOPT and SBB as well as the global optimizers SCIP 3 and BARON 12. All the input data are directly or indirectly derived from literature (Maenza and Posti, 2013).

In this case study, we consider a network with 2 freshwater sources, 3 shale sites, and 10 wells in each site. Freshwater withdrawal availability is estimated based on historical data with a consideration of seasonal fluctuation (Yang and Grossmann, 2013). There are 3 different CWT facilities and 10 available disposal wells. Wastewater is classified into 3 different TDS concentration ranges based on the TDS concentration constraints of different treatment technologies, ranging from 0 to 20,000 mg/L, 20,000 to 40,000 mg/L, and greater than 40,000 mg/L, respectively (Gaudlip and Paugh, 2008). We note that for Marcellus shale play, available disposal wells are mainly located far away from the shale sites in Pennsylvania (e.g. in Ohio), which are not preferable due to the high transportation cost. There are 3 levels of onsite treatments. To be more specific, we consider the sequential treatment processes including coagulation, flocculation, and disinfection steps for the primary treatment; the application of hydrated lime (Ca(OH)2) and corresponding clarifier and filter facilities for the secondary treatment; and thermal distillation technology for tertiary treatment. We consider pipeline and truck transportation modes for freshwater transportation from freshwater sources to shale sites. Truck is considered as the only transportation mode for wastewater from shale sites to CWT facilities and disposal wells. We also consider 3 capacity ranges for the pipelines used for transporting freshwater and 3 capacity ranges for each level of onsite treatment facilities. The planning horizon is 10 years.

	Discrete	Continuous	Constraints	Objective	Solution time
	variables	variables	Constraints	value	(CPUs)
MILFP problem (P)					
Parametric	297	158,213	208,080	14,970	27
R-L	297	158,510	208.970	14,970	294
B&B + C-C	0 ^a	158,213	208,080	14,970	11,272
DICOPT	297	158,213	208,080	14,970	7,180
SBB	297	158,213	208,080	14,970	29,486
SCIP 3	297	158,213	208,080	N/A ^b	N/A ^b
BARON 12	297	158,213	208,080	N/A ^b	N/A ^b
MILP problem (EP)					
CPLEX 12.6	297	158,213	208,080	4,533,300	5

Table 2. Summary of model statistics and computational results for case study	Table 2.	Summary	of model	statistics and	l computational	results for	case study	1
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a. For each iteration after reformulation and relaxation, there are no discrete variables.

b. Solver failure encountered.

As can be seen in Table 2, all of these solution methods return the same optimal solution to the MILFP problem (P), except SCIP 3 and BARON 12. The parametric algorithm based on the exact Newton's method is the most efficient algorithm. It takes only three iterations and a total of 27 CPUs to converge to the global optimal solution. The reformulation-linearization algorithm also has excellent computational performance with a computational time of 294 CPUs. The Branch-and-Bound & Charnes-Cooper transformation algorithm returns the same objective value as the other approaches in 11,272 CPUs with a total of 183 nodes fathomed in the Branch-and-Bound tree. For the MILP problem (EP), the objective of which is to minimize the total cost, we directly apply the MILP solver CPLEX 12.6 to solve the problem with only 5 CPUs.



Figure 4. Water supply chain network comparison between (a) MILFP problem (P) and (b) MILP problem (EP)

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As can be seen in Figure 4, in problem (P), all the three-level onsite treatments are applied, while in (EP), tertiary treatment is not chosen. Therefore, more wastewater is transported to the CWT for treatment and discharge. As to the freshwater acquisition, in problem (P), a total of 5,268,306 barrels of freshwater is withdrawn from freshwater source 1 over the 10-year time horizon. In contrast, 5,912,385 barrels of freshwater is withdrawn in (EP), which is 12% more than that in problem (P), resulting in 11% higher water supply cost. Obviously, the freshwater consumption and the corresponding cost are significantly reduced in problem (P). Detailed water management strategy is given in Figure 5. The water management strategies for both problem (P) and (EP) share some common features: disposal wells are not employed due to the long-distance transportation cost; CWT as well as onsite treatment options are applied to treat the wastewater; the treatment load for onsite treatment is relatively stable, while the amount of wastewater treated by CWT fluctuates with time; storage option behaves as a "buffer" to compensate the gap of wastewater treatment amount of CWT. The major difference between these two water management strategies is the preference of onsite treatment over the CWT. In problem (P), the amount of wastewater treated onsite and reused is almost twice as much as that in MILP problem (EP). As a result, less wastewater is transported to CWT facilities for treatment, and the inventory level is lower in the MILFP problem (P). Consequently, less water is reused onsite or recycled from CWT facilities, and more freshwater is required at each time period in the MILP problem (EP).



Figure 5. Water management strategy comparison between (a) MILFP problem (P) and (b) MILP problem (EP)

Table 3. Cost breakdown of water management options in MILFP problem (P) and MILP problem (F	problem (P) and MILP problem (EP)
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	Water acquisition	Transportation	Onsite treatment	CWT	Storage
MILFP problem (P)	3%	6%	79%	11%	1%
MILP problem (EP)	4%	29%	13%	51%	3%

The overall cost distribution of different water management sections is given by Table 3. The profit gained corresponding to unit freshwater consumption is \$14,970 per thousand barrels of freshwater consumption in problem (P), 14% greater than \$13,084 per thousand barrels of freshwater consumption in problem (EP), indicating a higher freshwater utilization efficiency. As to the detailed breakdown of total water management cost, the CWT facilities contribute 11% of the overall cost for water management in problem (P), while in (EP), this number is 51%; onsite treatment accounts for 79% of the total cost in problem (P), while in problem (EP) only 13% of the total cost comes from onsite treatment. Due to the extensive application of onsite treatment and reuse in MILFP problem (P), the stress on freshwater withdrawal, related transportation, and onsite storage is relieved. As a result, the costs of freshwater acquisition, storage, and transportation in problem (P).

Based on the comparison above, we conclude that it is impossible to obtain a good balance between cost effectiveness and freshwater conservancy by simply minimizing the total cost. In the MILP problem (EP), most of the wastewater is transported to the CWT for treatment and direct discharge. In contrast, the fractional objective function in problem (P) gives a more efficient and practical water management strategy with more water reuse, less freshwater consumption, and less stress on transportation and storage. Such a water management strategy is close to the real one applied at Marcellus shale play (Wilson and VanBriesen, 2012).

5. Conclusions

In this paper, we presented a novel optimization model for the optimal design and operations of water supply chain networks for shale gas production. The objective function was given as the profit per unit freshwater consumption. Such a problem was formulated as an MILFP problem. While the models were general enough to consider multiple water management options, we focused on the disposal, CWT, and onsite treatment options in this work. Furthermore, three levels of onsite treatment were considered. To illustrate the proposed models and solution methods, one case study was presented based on Marcellus shale play. The optimal water management strategies obtained from MILFP problem (P) and MILP problem (EP) were carefully analyzed and discussed. The superiority of problem (P) over problem (EP) was validated. Moreover, the onsite treatment option turned out to be appealing in improving freshwater conservancy, maintaining a stable water flow, and reducing transportation burden. The computational results indicated the outstanding efficiency of the parametric algorithm in solving large-scale MILFP problems.

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